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Synthesis of α-β Bi<sub>2</sub>O<sub>3</sub> heterojunction photocatalyst and evaluation of reaction mechanism for degradation of RhB dye under natural sunlight

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Dear Sir/Madam,

Please find the revision of the revised manuscript CERI-D-19-05645 (ms) entitled, 'Synthesis of α-β Bi2O3 heterojunction photocatalyst and evaluation of reaction mechanism for degradation of RhB dye under natural sunlight' by Kamal Kanti Bera, Malay Chakraborty, Mousumi Mondal, Senjuti Banik, Swapan Kumar Bhattacharya\* for your kind perusal and onward publication. The revision has been made taking consideration of all suggestions of reviewer # 2 and where we feel it necessary. The changes made are highlighted in green colour. Response to the reviewers' comments has also been attached herewith.

Hope, you will find the manuscript worthy for publication in your esteemed journal considering its importance.

Hope you will be kind enough to do the needful for publication of the paper at the earliest.

Thanking you,

Dated: 20.11.2019 Yours truly,

Swapan Kumar Bhattacharya

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# **Synthesis of α-β Bi2O3 heterojunction photocatalyst and evaluation of reaction mechanism for degradation of RhB dye under natural sunlight**

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# **GRAPHICAL ABSTRACT**



#### **Abstract**

A few biphasic nano composites containing  $\alpha$  and  $\beta$  Bi<sub>2</sub>O<sub>3</sub> of varying composition were synthesized by facile solvothermal method without using any capping agent and further calcination. X-ray diffraction, microscopic and spectroscopic techniques were employed for characterization of the as synthesized catalysts which are used as photocatalysts in degradation of pollutant, Rhodamine B (RhB) dye. The band gap of the nanocatalysts as calculated from tauc plot varies within 2.35- 2.58 eV for β-form and 2.85-3.19 eV for α-form in the α-β Bi<sub>2</sub>O<sub>3</sub> hetrojunctions. The operational parameters that influence the degradation process were optimized. The best catalyst dosage and pH are  $0.5 \text{ gL}^{-1}$  and 4 respectively and the best concentration of  $H_2O_2$  when added is 2 mM for 10 ppm aqueous solution of dye. Among different heterojunctions, the best catalyst which is produced from bismuth nitrate concentration of 0.05 M, degrades RhB upto 99.6% at pH 4 under 120 min sunlight irradiation. The effects of addition inorganic salts in RhB dye solution were also examined. The radical trapping experiments have been applied to explore the involved and main species responsible for degradation. The identification of degradation products of RhB was analyzed and the plausible mechanistic pathway is drawn from HPLC and HRMS. It shows that the degradation of RhB proceeds via initial generation of N-deethylated products followed by ring opening ones, which indicates the photosensitization induced photocatalytic mechanism of the reaction.

**Keywords:** α-β Bi<sub>2</sub>O<sub>3</sub> hetrojunction; Solar remediation photo catalytic activity; Rhodamin B degradation products; Reaction mechanism

# **1. Introduction**

In heterogeneous photocatalysis, advanced oxidation process (AOP) has been extensively used as an eco-friendly, cost-effective and green process for productive degradation of various harmful organic pollutants [1, 2] and provides solution to water quality problems. Among the various AOP, coupling of two semiconductors with different band gaps is a well-advocated strategy which increases the absorption and decreases the electron-hole recombination process. Semiconductors like TiO<sub>2</sub>, ZnO, Bi<sub>2</sub>O<sub>3</sub>, NiO [3-6] etc are usually used in AOPs. Extensive research was done on  $TiO<sub>2</sub>$  due to its high abundance, low cost and good stability. However high band gap  $(E_g 3.2eV)$  and poor absorption of sunlight  $(4-5%)$  [7] limit its performance. Therefore visible light active photocatalyst,  $Bi<sub>2</sub>O<sub>3</sub>$  having lower band gap (2.4-2.9eV) is chosen [8] as an alternative. Among six main crystallographic polymorphs, α (monoclinic), β (tetragonal), γ (cubic bcc), δ (cubic face centred), ε (orthorhombic), ώ (triclinic), the first one is stable at low temperature, the last one at high temperature and others are metastable state [9]. A few researches on photocatalytic activity of  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> heterojunctions are summarized in Table1 [10-17] and illustrate their importance. The table shows that the degradation efficiency of the toxic and carcinogenic dye, Rhodamin B (RhB) in natural sunlight was presented by only two previous works [12, 15]. In our previous work [12], the dependence of synthesis temperature on the composition  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> nano heterojunction and synergistic effect on degradation of RhB over  $\alpha$ or β Bi<sub>2</sub>O<sub>3</sub> in sunlight were studied and the maximum efficiency was found to be 99.7%. In the present paper, we synthesize different  $\alpha$ - $\beta$  Bi<sub>2</sub>O<sub>3</sub> heterojunctions by varying molar concentration of the precursor bismuth nitrate, studied their photocatalytic activity, optimize some degradation

**Table 1:** Summary of recent relevant works, synthesis and photocatalytic degradation of various dye on α-β $\rm{Bi}_2\rm{O}_3$  heterojunction catalyst.





parameters like pH, effective catalyst dosage in details and the effect of addition of  $H_2O_2$  and inorganic salt like Cl, NO<sub>3</sub>, SO<sub>4</sub>, CH<sub>3</sub>COO in sunlight. The degradation intermediates of RhB by  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> photocatalyst in natural sunlight were also identified with assistance of high resolution mass spectroscopy (HR-MS) and high performance liquid chromatography (HPLC) study.

# **2. Methods**

# *2.1 Synthesis of Bi2O3 nano crystal*

All chemicals were analytical reagent grade and used without purification. In this experiment, α-

 $β$  Bi<sub>2</sub>O<sub>3</sub> heterojunctions were synthesized by a facile hydrothermal method. In a typical synthesis, 0.02 (M) Bismuth nitrate pentahydrate  $[Bi(NO<sub>3</sub>)<sub>3</sub>, 5H<sub>2</sub>O]$  was dissolved at first in 50 mL glycerol with sonication for 1 h followed by constant stirring for 3 h. Then it was transferred into a 100 mL teflon-lined stainless steel autoclave, sealed and heated at  $180^{\circ}$ C for 4 h. The system was then cooled at room temperature and the product obtained by filtration, washed several times with distilled water and ethanol and finally dried in vacuum desiccators at room temperature. Other catalysts were prepared by changing the precursor concentration from 0.02 (M) to 0.3 (M) keeping other procedure same. The precursor concentrations are taken 0.02 (M), 0.03 (M), 0.05 (M) and 0.3 (M) and the corresponding products are coded as S1, S2, S3 and S4. The reaction of glycerol with bismuth nitrate is shown in the Scheme 1[14].



**Scheme 1.** Synthesis of  $Bi<sub>2</sub>O<sub>3</sub>$ .

#### *2.2 Structural characterization*

X ray powder diffraction (XRD) study was carried out using a difractomter (Bruker D8 Advance) equipped with a CuK $\alpha$  radiation source ( $\lambda$  = 1.5418A<sup>°</sup> generated at 40 KV and 40 mA). The morphologies of obtained materials were inspected by a field emission scanning electron microscopy (FESEM, INSPECT F50 FEI), transmission electron microscopy (TEM), highresolution transmission electron microscopy (JEM-2100 HRTEM, JEOL, Japan, operated at

200Kv), selected-area electron diffraction (SAED) and energy dispersion X ray spectroscopy (EDX). The ex- situ Fourier transform infra-red spectroscopy (FTIR) of synthesized materials were carried out using spectrophotometer (Perkin Elmer, SN-74514, spectrum RX1 and resolution 4 cm<sup>-1</sup>). Absorption spectra were recorded using a model V-630 UV-visible spectrophotometer (Jasco, Japan) within the wavelength range of 400-600 nm. Nitrogen adsorption and desorption experiment was carried at 77 K in a NOVA 2200e, Quantachrome Instruments, USA, specific surface area and pore size analyzer. In this measurement, the samples were first degassed under vacuum condition for 8 h at 423 K. The surface charge of  $Bi<sub>2</sub>O<sub>3</sub>$  was characterized by Zeta potential analyzer (Zetasizer nano ZS (Malvern,UK) at 90◦ scattering angle with a He-Ne Laser( $\lambda$  = 632.8nm)

# *2.3 Photocatalytic experiments*

The photocatalytic activities of the synthesized  $Bi<sub>2</sub>O<sub>3</sub>$  nano particles were carried out by photo degradation of Rhodamine-B under sunlight. 100 mL of 10 ppm of aqueous suspension of Rhodamine B containing  $0.5 \text{ gL}^{-1}$  of Bi<sub>2</sub>O<sub>3</sub> nano particles agitated by a magnetic stirrer (500) rpm) in the dark to make the dye- adsorbed catalyst in the well dispersed state in the solution. Then it was exposed to sunlight and 4 mL solution was withdrawn at every 15 min interval, centrifuged and filtered. The residual dye solution was analyzed spectrophotometerically and the changes in the concentration of Rhodamine-B were observed from its characteristics absorption at the wavelength of absorption maximum, 554 nm. To keep the consistency in irradiation of natural sunlight all the photocatalytic experiments were intentionally performed between 11.00 a.m. to 2.30 p.m. in consecutive sunny days during February and March, 2018 (Jadavpur, Kolkata, India). Besides a digital lux meter (LX 101A.) is used to measure time dependent light intensity of natural sunlight from which average illumination is obtained. The photo catalytic degradation was calculated as follows:

% Photocatalytic degradation =  $\{ (C_0 - C_t) / C_0 \}$  x 100

Where  $C_0$  is the initial absorbance and  $C_t$  the absorbance of the sample after irradiation for time t minute.

#### *2.4. Trapping Experiment*

The method of trapping experiment is the same to the photocatalytic degradation experiments; only difference is the addition of different scavengers. In this experiment we used 2 mM of different scavengers like Parabenzoquinone (BQ), Isopropyl alcohol (IPA) and Sodium-EDTA  $(Na<sub>2</sub>EDTA)$  as  $O<sub>2</sub>$ ,  $OH$  and  $h<sup>+</sup>$  trappers respectively.

#### **3. Result and discussions**

# *3.1 Characteristics of Bi2O3 nano particles*

Fig. 1 illustrates the XRD patterns of as-prepared samples. The sample S1, S2, S3 exhibits peaks having 2θ (/degree) values at 28.00, 30.18, 32.90, 42.16, 53.40, 62.93, 68.56, 78.35 corresponding to (201), (211), (220),(320), (322), (422), (432), (314) planes of tetragonal β-Bi<sub>2</sub>O<sub>3</sub> respectively [JCPDS NO 78-1793] and the peaks having  $2\theta$ (/degree) at 23.92, 29.36, 35.09, 47.33, 54.67, 56.56 can be attributed to  $(1\bar{1}1)$ ,  $(311)$ ,  $(210)$ ,  $(420)$ ,  $(014)$ ,  $(240)$  planes of monoclinic  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> respectively [JCPDS NO 72-0398]. XRD pattern recorded for sample S4 contains peaks at 2θ (/degree) values of 30.18, 32.90, 39.72, 42.16, 53.22, 62.23, 64.63 corresponding to the (211), (220), (320), (402), (322), (422), (520) planes of tetragonal β-Bi<sub>2</sub>O<sub>3</sub> [JCPDS NO 78-1793] and peaks at 2θ values of 23.92, 27.12, 37.83, 44.71, 46.13, 48.71, 56.12, 70.72 for ( $1\overline{1}1$ ), (112), ( $1\overline{1}2$ ), (510), (040), (041), (104), , (061) planes of monoclinic  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> [JCPDS NO 72-0398] respectively. The size of crystallites of the samples was measured using the Debye-Sherrers equation [18]. From XRD data we concluded that our synthesized samples were the mixtures of  $\alpha$  and  $\beta$  phases of Bi<sub>2</sub>O<sub>3</sub> with various compositions and with crystallite size in the range of 6-42 nm. The composition of heterojunction with different  $\beta/\alpha$  phase is also a function of molar concentration of the precursor. Table 3 presents the relative ratio of intensity of one dominant peak of  $\beta$  at  $2\theta = 32.90^{\circ}$  (corresponding to 220 plane) to another significant peak intensity of  $\alpha$  at  $2\theta = 27.10^{\circ}$  (corresponding to 112 plane) for different synthesized samples. The data reveal that the relative increase of β over α phase continues up to concentration limit of 50 mM of the precursor. At higher concentration of  $Bi(NO<sub>3</sub>)<sub>3</sub>$  like 300 mM, the heat rejection is relatively slow during cooling of autoclave and thus the  $\alpha$  form which is more stable at higher temperature predominates. Notably, pure  $\alpha$  and  $\beta$  form can be obtained by the calcination of  $\alpha$ -β  $Bi<sub>2</sub>O<sub>3</sub>$  at 400°C and 250°C respectively in hydrothermal method [12]. Thus the relative composition of  $\alpha$  and  $\beta$  in  $\alpha-\beta$  Bi<sub>2</sub>O<sub>3</sub> heterojunctions can be monitored by changing the concentration of the precursor in the hydrothermal synthesis.



**Fig. 1.** The XRD patterns of four (S1-S4)  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> heterojunctions synthesized at different concentration of precursor.

Fig. 2 a, b and c. demonstrate the FESEM images of S1, S3 and S4  $\alpha$ - $\beta$  Bi<sub>2</sub>O<sub>3</sub> composites respectively. From Fig. 2 it is seen that the S1 catalyst contains the mixture of nano sphere of diameter 24 nm and nano rod like structures with average 126 nm lengths and 40 nm breadths. S3 catalyst displays nano flake like structures which are made by spherical particles of average diameter 28 nm. S4 catalyst shows the budding flower like morphology. The representative EDS spectrum of some parts of S1 (Fig. 2d) confirms the formation of  $Bi<sub>2</sub>O<sub>3</sub>$  in the photo catalysts as Bi and O atom are present in the ratio of 2: 3.06.



**Fig. 2.** The FESEM images of (a) S1 (b) S3 (c) S4 photocatalyst (d) EDX image of representative  $α$ -β Bi<sub>2</sub>O<sub>3</sub> heterojunction.

The TEM images of S3 catalyst (Fig. 3a and b) indicate that the catalyst is composed of porous nano flakes of  $Bi_2O_3$ . The average size of the pores as determined by bar diagram is about 4.04 nm. The HRTEM image presented in figure 3c displays the lattice spacing of 0.31 nm and 0.37 nm corresponding to the inter planner spacing of (201) of β and  $(111)$  of α Bi<sub>2</sub>O<sub>3</sub> which further confirm a phase junction between  $\alpha$  Bi<sub>2</sub>O<sub>3</sub> and  $\beta$  Bi<sub>2</sub>O<sub>3</sub>. Selected-area electron diffraction (SAED) profile (Fig. 3d) exhibits also  $\alpha$  and  $\beta$  planes matched with the XRD results.

 $N_2$  adsorption- desorption isotherm measurements were carried out to evaluate the specific surface areas, pore volumes and the average pore size distributions of the as prepared samples. The BET surface area  $(m^2g^{-1})$  of S1-4 samples was found to be 19.6, 11.5, 49.1 and 4.24 and Barrett-Joyner-Halenda (BJH) desorption analyses for the composites exhibit that the least value of average pore-diameter (2.68 nm) is obtained with S3 among all the synthesized materials. The highest surface area (11.5 times than S4 catalyst), the largest total pore volume (0.108 cm<sup>3</sup>g<sup>-1</sup>) and the smallest pore diameter (2.68 nm) of S3 catalyst (Table 2) could supply more surface active sites and make transport easier leading to an enhancement of photocatalytic performance. According to the Brunauer-Deming-Deming-Teller (BDDT) classification, the isotherms ( Fig.4a) can be categorized as Type IV hysteresis loop indicating the formation of mesoporous materials [16, 19-20]. Fig.4b provides the Pore size distributions of as-synthesized  $\alpha$ - $\beta$  Bi<sub>2</sub>O<sub>3</sub> heterojunctions (S1-S4).



Fig. 3. (a) The TEM image at a given magnification and (b) same image at higher magnification (c) HRTEM image (d) SAED pattern of S3 catalyst.



Fig. 4. (a) N<sub>2</sub> adsorption-desorption isotherms (b) Pore size distributions of as-synthesized  $\alpha$ - $\beta$ Bi2O3 heterojunctions (S1-S4).

Table 2. BET surface areas, total pore volume and mean pore size of synthesized catalysts.



Fig. 5 illustrates the FTIR spectra of different  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> heterojunctions (S1-4). The band at 850 cm<sup>-1</sup> is attributed to the symmetrical stretching vibration Bi-O bond of the BiO<sub>3</sub> species [12, 21] and the band 768 cm<sup>-1</sup> originates from the metal oxygen vibration [22]. The band at 1558 cm<sup>-1</sup> is related to the vibration of the residual hydroxyl group  $(Bi<sub>2</sub>O<sub>3</sub>–OH)$  [23] and the weak broad band centered around  $3322 \text{ cm}^{-1}$  is possibly due to the stretching vibration of O-H bond of very small number of adsorbed H<sub>2</sub>O molecules [24, 25]. The band at 1386, 1287 and 1459 cm<sup>-1</sup>can be related to the C-O stretching vibrations and –C-H bending vibration of the adsorbed ethanol [26, 27]. The band at 921 cm<sup>-1</sup> is the characteristic stretching vibration of Bi-O bond in strongly distorted  $BiO<sub>6</sub>$  octahedral units [28].



**Fig. 5.** FTIR spectra of synthesized  $\alpha/\beta$  Bi<sub>2</sub>O<sub>3</sub> nano hetero-junctions (S1- S4).

The optical properties of synthesized  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> are followed by UV-vis absorption spectroscopy.

The order of absorbance of the heterojunctions is  $S3 > S1 > S2 > S4$  where corresponding profiles are exhibited in the Fig. 6a. Table 3 reveals that S1- S2- S4 size increases, surface area decreases and absorbance decreases. Greater absorbance of S3 is due to high β content. In earlier report we find that the absorbance of β  $Bi<sub>2</sub>O<sub>3</sub>$  is enhanced in the visible light region compared to  $\alpha$  Bi<sub>2</sub>O<sub>3</sub> [29]. The band gap energy (E<sub>g</sub>) was evaluated using Tauc plot as done in our previous work [12]. The plot of  $(ahv)^2$  vs hv for each composite exhibits two different slopes of the tangents from which the bands of  $\alpha$  and  $\beta$  phases of Bi<sub>2</sub>O<sub>3</sub> are obtained as depicted in Fig. 6b and Table 3 respectively. The band gap varies 2.35-2.58 and 2.85- 3.19 for  $\beta$  and  $\alpha$  phases in  $\alpha$ - $\beta$  $Bi<sub>2</sub>O<sub>3</sub>$  heterojunctions respectively. The obtained band gap values are consistent with the previous results [12, 16, 25- 31].

**(a) (b)** 



Fig. 6. (a) UV-vis spectra of S1-S4 photo-catalysts. (b)  $(\alpha h v)^2$  vs hv profiles of S1-S4 photocatalysts for the determination of band gap energy.

#### *3.2. Influence of reactant concentration*

The size of synthesized  $Bi<sub>2</sub>O<sub>3</sub>$  nano particles has also great dependence on the concentration of  $Bi(NO<sub>3</sub>)<sub>3</sub>$  solution in the hydrothermal synthesis. Fig. 7 represents the variation of particle size with concentration of reactant. From the Fig. 7 it is seen that average particle size gradually increases from 6 nm to 42 nm when concentration of reactant changes from .02 (M) to 0.3 (M). In general, when concentration the reactant was low, the synthesized particles are accumulated in different localized zone and agglomerate as glycerol is highly viscous [32]. As the ratio bismuth nitrate: glycerol increased, accumulation of particles was greater causing agglomeration and higher particle size.



**Fig. 7.** Effect of precursor concentration on the average size of the synthesized particles which are obtained by heating hydrothermally at temperature 180◦C for 6 h, all other parameters being unchanged.

# **4. Photocatalytic experiment of Rhodamine-B**

#### *4.1 Effect of irradiation time*

To evaluate the photocatalytic activity of synthesized  $\alpha$ - $\beta$  Bi<sub>2</sub>O<sub>3</sub> photocatalysts, the degradation of RhB are performed upto 180 min under sunlight. The degradation profiles of S1, S3 and S4 catalysts are exhibited in Fig. 8 a, b and c respectively. The plot of  $c/c<sub>0</sub>$  versus time in Fig. 8d exhibits 84.4, 80.2, 89 and 31.2 % degradation of RhB and associated rate constant of the reaction in min-1, are 0.194, 0.133, 0.02037 and 0.00201 for S1, S2, S3 and S4 catalysts respectively.







**Fig. 8.** UV-vis spectra of RhB degradation at different time intervals using (a) S1 (b) S3 (c) S4 catalyst and (d) Plot C/Co versus time for pure dye and S1-S4 catalysts [RhB]  $=1\times10^{-5}$  M, [catalyst] = 0.5 g/L. and AI = 490×100 lux.

The degradation rate is the highest for S3 catalyst and 10 times greater than S4 catalyst. This is due to the high β/α ratio, lower size and lower band gap in S3 catalyst compared to S4. The degradation rate also slightly decreased for S2 compared to S1 catalyst because of greater size of the later. From Fig. 8a and b, it is shown that the spectral maximum shifted to 554 to 496 nm gradually after 60 min irradiation in sunlight and the color of the solution changed from red to colorless. The decrease in absorbance and hypsochromic shift in absorption spectra is due to breaking down of RhB structure and step by step de-ethylation respectively [33, 34]. The influence of degradation parameters like catalyst dosage, pH, addition of  $H_2O_2$  and inorganic salt in presence of  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub>(S3) catalyst are discussed below.

#### *4.2 Effect of solution pH*

The degradation was studied in the pH range 2-12 (Fig. 9a) and adjusted by addition of strong HCl and NaOH solutions. The figure shows that S3 catalyst  $(0.2 \text{ gL}^{-1})$  degrades RhB dye (10) ppm) upto 94% in 60 minute at pH 4. The rate constant  $(x10^2)$  values  $(min^{-1})$  are 3.26, 6.28, 5.14, 2.84, 1.71 and 1.67 at pH 2, 4, 6, 8, 10 and 12 respectively. It reveals that the degradation is very poor at both very low and high pH. At very low pH, increased concentration of Cl might act as a scavenger of h<sup>+</sup> and **OH** radical [10]. At very high pH, the holes may be absorbed by OH ions producing OH radical which otherwise engages to produce  $H_2O_2$  and decomposition of dye [12]. The rate constant values exhibit a maximum at pH 4 as evident from the lower inset of figure 9a. The surface charge of the  $Bi<sub>2</sub>O<sub>3</sub>$  nano particles was determined by Zeta potential measurement. For its measurement  $Bi_2O_3$  suspensions (0.1 gL<sup>-1</sup>) were prepared and the pH was adjusted between 1-10 using 1 mol Lit<sup>-1</sup> NaOH or HCl solutions [35]. The Zeta potential (ζ) of Bi<sub>2</sub>O<sub>3</sub> catalyst at different pH conditions was measured to determine the effect of pH values on the surface charge of  $Bi_2O_3$ . The change of  $\zeta$  of S3 catalyst (Fig. 9b) shows an evident decrease with increasing pH values. At pH 2, value of ζ is 8.17 mV and pH 3 it becomes -9.41mV. The point of zero charge (PZC) of the  $Bi_2O_3$  is estimated to be 2.45. At pH (4) > PZC (2.45), surface of  $Bi_2O_3$ becomes negatively charged. Therefore, -vely charged  $Bi<sub>2</sub>O<sub>3</sub>$  particles facilates the adsorption of + vely charged diethylamine group in RhB molecules owing to electrostatic attraction and more adsorption of RhB dye on the COO part of zwitter ionic RhB dye (since its pKa is 3.1). This also infers that degradation of RhB undergoes via N-deethylation mechanism predominantly [36].



**Fig. 9** (a) effect of solution pH on chronodegradation of RhB. Initial concentration and catalyst are  $1x10^{-5}$  M and  $0.2gL^{-1}$ . (b) Effect of pH on the zeta potential of S3 composite in aqueous solution ( $T = 25°$  C). (c) Effect of dosages of S3 catalyst on chronodegradation of RhB upto 60 min, initial concentration of which is  $10^{-5}$  M at pH 4. The variation of rate constant with pH and catalyst dosages is presented at the lower inset of each figure. The average illumination (AI) of the natural sunlight during the experiment was 630 x100 lux. (d) Degradation profile of RhB in presence or absence of H<sub>2</sub>O<sub>2</sub> [dye] =1x10<sup>-5</sup> M, [catalyst] = 0.5 gL<sup>-1</sup>, pH=4, AI= 480x100 lux., irradiation time 120 minutes under sunlight. (e) Effect of variation of  $H_2O_2$  concentration on degradation .The lower inset represents apparent rate constant of degradation versus  $H_2O_2$ concentration in the solution. (f) Plot C/Co versus time for the addition of inorganic salt in the photo catalytic degradation of RhB.

### *4.3 Effect of catalyst dosage*

The optimum value of catalyst dosage obtained by studying with various concentration of S3 catalyst from 0.05  $gL^{-1}$  to 1.5  $gL^{-1}$  taking 10 ppm RhB solution at pH 4 (Fig. 9c). The degradation is achieved by 98% with 0.5 gL<sup>-1</sup> in 60 min. The rate constant (x100) values (min<sup>-1</sup>) are 0.387, 4.44, 5.99, 6.27, 3.23 and 0.578 for 0.05, 0.1, 0.2, 0.5, 1, and 1.5  $gL^{-1}$  respectively. The variation of rate constant with catalyst dosage is presented at the lower inset of figure 9c. It exhibits initial increase of degradation efficiency plausibly because of increasing number of reaction sites on the catalyst surface. Again it decreases with higher loading is due to the light scattering effects and self quenching of excited catalyst molecules [37-40].

*4.4 The effect of addition of H2O2*

The effect of addition of  $H_2O_2$  on the degradation of RhB (Fig. 9d) in presence of catalyst under natural sunlight was also studied. When 1 mL of 2 mM  $H<sub>2</sub>O<sub>2</sub>$  (30%) is added to 100 mL of 10 ppm dye solution, it degrades 99.6% and 89% in presence and absence of  $H_2O_2$  respectively in 120 min.  $H_2O_2$  increases the rate of degradation of RhB by 10.6%. The rate constant values are 2.1x10<sup>-4</sup>, 1.1x10<sup>-3</sup>, 2x10<sup>-2</sup>, and 4.4x10<sup>-2</sup> min<sup>-1</sup> for pure dye, dye with H<sub>2</sub>O<sub>2</sub>, dye with catalyst and dye with both catalyst and  $H_2O_2$  respectively. The rate constant value of the dye in presence of catalyst increases 2.16 times on addition of  $H_2O_2$  to it.

In order to investigate the effect of  $H_2O_2$  concentration on the degradation kinetics, different quantity of  $H_2O_2$  is added to the solution keeping other parameters constant. Fig. 9e reveals that the rate constant of the decolonization reaction increases by 32.7% in 60 min with increasing concentration of  $H_2O_2$  from 0.15 mM to 2 mM. This might be due to increase of  $\overline{O}H$  radical concentration on increasing the concentration of  $H_2O_2$  in the solution.

The increase of  $H_2O_2$  concentration further results in decrease of removal of color by 20.4%. Therefore the concentration of  $H_2O_2$  kept at an optimum level (2 mM). This might be due to the fact that at high concentration very reactive 'OH radical could be consumed by H<sub>2</sub>O<sub>2</sub> and results in the generation of less reactive **.** OOH radical which has lower oxidation capabilities than **.**OH [41] and as be illustrated by the following Eqs.(1)- (3).





$$
HO_2^{\bullet} + O_1 \longrightarrow H_2O + O_2 \tag{3}
$$

*4.5 Effect of inorganic salts*

The photo-catalytic degradation may be affected by many organic and inorganic species present in the real industrial waste water. In this study, the effect of  $NO<sub>3</sub>$ , Cl,  $SO<sub>4</sub><sup>2</sup>$ , CH<sub>3</sub>COO<sup>-</sup> ions on RhB dye was investigated and depicted in Fig. 9f. From the figure, it is seen that the % of degradation of RhB in presence of NO<sub>3</sub>, Cl,  $SO_4^2$ , CH<sub>3</sub>COO<sup>-</sup> ions are 96.7, 93.6, 81 and 22% respectively. Notably, % of degradation in absence of any such ions is 88% in 90 min. The higher degradation efficiency in presence of  $NO<sub>3</sub>$  ion is due to strong adsorption of nitrate ions on α-β Bi<sub>2</sub>O<sub>3</sub> surface and production of OH radicals from photo induced aqueous nitrate in sunlight  $[42-44]$  as shown in the Eq.  $(4)$  -  $(7)$ .

$$
NO_3 + hv \longrightarrow NO_2 + O \tag{4}
$$

$$
O^+ + H_2O \longrightarrow OH + OH^-(5)
$$

$$
NO_3^- + hv \longrightarrow NO_2^- + O(^3P) \tag{6}
$$

$$
NO_2 + O(^3P) + hv \longrightarrow NO + OH \tag{7}
$$

It has been illustrated in our previous result [12] that the introduction of Cl into the reaction mixture of degradation causes modification of  $Bi<sub>2</sub>O<sub>3</sub>$  surface by producing BiOCl which can also assist the degradation process [11] but Cl- ion also acts as **.**OH radical scavenger by the following Eq.  $(8)$  -  $(11)$  [45, 46]. So, the rate in presence of Cl decreases little bit compared to that in presence of  $NO<sub>3</sub>$ .

$$
Cl^+ + h^+ \longrightarrow Cl^+ \tag{8}
$$

$$
CI^+ OH \longrightarrow COH^-
$$
 (9)  
2 CI
$$
CI_2 + OH \longrightarrow COH^-
$$
 (10)

$$
Cl2 + H2O \longrightarrow HOCI + CI + H+
$$
 (11)

For  $SO_4^2$  ions, they are adsorbed in the surface of the  $Bi_2O_3$  and reacts with photo induced holes  $(h<sup>+</sup>)$  to form  $SO<sub>4</sub>$  [47, 48]. This generated sulfate radical also helps the degradation process according to the following Eqs. (12) and (13) but it also acts as a holes scavenger. So, degradation rate decreases compared to nitrate and chloride ion.

$$
h^+ + SO_4{}^{2-} \longrightarrow SO_4{}^{2-} \tag{12}
$$

$$
SO_4^{-+} + e^{-} \text{ (or H2O)} \longrightarrow SO_4^{2-} \text{ (or } SO_4^{2-} + OH + H^+)
$$
 (13)

But the approach of  $CH<sub>3</sub>COO<sup>-</sup>$  ion demonstrated the adverse effect. Due to addition of CH3COONa salt in the solution, the medium becomes basic and degradation decreases which is discussed in earlier. It also inhibits the formation of OH by the photo-kolbe reaction [49, 50] via the Eqs. (14)-(16).

### 4.6 Effect of Scavengers

In order to explore the different species involved and the main agent responsible for the photocatalytic degradation of RhB, the reaction is followed in presence and absence of scavenger of h<sup>+</sup>, 'O<sub>2</sub>' and 'OH radicals. In the trapping experiments each scavenger was added at equal concentration in separate experimental solutions containing same composition of dye,  $H_3O^+$  and catalyst to observe the direct influence of the scavengers on the degradation rates. As shown in the Fig. 10, the degradation efficiency of RhB is found to decrease rapidly from 98.66% to 19.85% with addition of Na<sub>2</sub>EDTA which is a scavenger (trapper) of hole. The rate of arresting the reaction decreases in the order  $EDTA > BQ > IPA > No$  scavenger, indicating that the hole is the main species for degradation, followed by  $O_2$  and OH. The apparent rate constant (k) without addition of scavenger is  $0.07414$  min<sup>-1</sup> and for EDTA k values decreases to  $0.00072$  min<sup>-1</sup> <sup>1</sup>. A little decrease of rate constant for BQ (0.0565) and IPA (0.0682) is observed indicating h<sup>+</sup> is



the main reactive species and  $O_2$ . OH radicals plays minor role for the rapid degradation of RhB.

Fig. 10 Effect of different radical scavengers on the degradation of RhB over  $\alpha$ - $\beta$  Bi<sub>2</sub>O<sub>3</sub> under sunlight.

### *4.7 Identification of products of degradation of RhB with α-β Bi2O3 photocatalyst*

The RhB and intermediates generated in the photo-oxidative process were analyzed with a JASCO liquid chromatograph (HPLC) equipped a multi-wavelength detector and crestpak C18S column in the mixed eluent CH<sub>3</sub>OH:  $H_2O = 1:1$  by volume. The detected intermediates by HPLC are shown in the Figure 11a and b.

**Fig. (a)** 



Fig. 11. (a) Identification of the solute components in the solution undergoing photo catalytic degradation of RhB for different time extents by HPLC (under sunlight irradiation) (b) Some parts of profiles of figure 10 (a) (for the solution undergoes 150 min exposure of sunlight) with greater resolution to show the dominant products.

Five main components which have retention times (/min) 7.86 (m/z = 443, RhB), 5.82 (m/z = 387, DR), 4.68 (m/z = 387, EER), 3.43 (m/z = 359, ER) and 2.16 (m/z = 301) have been identified from the solution subjected to 90 min irradiation of sunlight. The absorbance (mAbs) of the peak obtained at retention time (/min )7.86 and 5.82 decrease and gradually lose m/z values of 28 and 45 which are the values of an ethyl group and one COOH group to form degraded products as presented in scheme 3. On expose of RhB solution to sunlight for 150 minute, destruction of xanthate ring occurs with formation of mainly adipic acid ( $Rt = 11.54$ ) min, m/z = 146), phthalic acid (Rt = 12.89 min, m/z = 166), 2- hydroxyl penta-di-ioic acid (Rt = 10.56 min,  $m/z = 148$ ) and 2, 5 di hydroxyl benzoic acid ( $R_t = 14.83$  min,  $m/z = 154$ ). All these products found from HPLC are well matched with the results of the other researcher [33, 51]. To obtain more information about the highly polar intermediates having small molar mass that could not be identified properly by HPLC, the products are further analyzed by HRMS technique. Some degradation products from HRMS, having strong sign at m/z 391, 387, 370, 301, 290, 197, 151, 149, 164 are presented in Fig. 12 and scheme 3. According to the relevant literature reported and combining with MS analysis, the peaks at m/z 443.4, 415.3, 387.3and 359.3 were identified as RhB and its N-deethylated intermediates namely, N,N,N'-triethyl rhodamine (DER), N, Ndiethylrhodamine (DR), N-ethyl-N-ethylrhodamine (EER) and N-ethyl rhodamine (ER) respectively. Among them, DR and EER are pair of isomers [52] and the compounds having m/z  $=$  370, 301 and 290 are the products of deethylation followed by de carboxylation. In our present study, HRMS provides the solid evidence for the formation N- dethylated products and cleavage of chromophore structure of RhB. Fig. 8a and b illustrates that absorption maxima of RhB dye shifted from 554 nm to 496 nm (hypsochromic shift). This shift of absorption band was presumed to result from the formation of a series of N-deethylated intermediates in a step wise manner [53, 54]. The color of the suspension changed gradually from pink to light green during the progress of the reaction and finally becomes colorless. This indicates that the  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> photocatalyst enhances significantly the de- ethylation process in comparison to the cleavage of ring structures [55].



**Fig. 12.** HRMS spectra of RhB solution under irradiation for (a) 0 min (b) 30 min (c) 60 min (d) 210 min.

# **Scheme 2.**

Photo catalytic degradation pathway of RhB by  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> photocatalyst.





Deethylation products

#### *4.8. Mechanism of photo degradation*

There are two possible photo chemical mechanisms: (a) photocatalytic process (b) photo sensitization process. In photo catalytic process, RhB can be degraded by direct interaction with strong oxidizing hole originating from the hybridization of the Bi 6s and 2p orbital. Our previous report [12] illustrates how the hole created by α-β  $Bi<sub>2</sub>O<sub>3</sub>$  photocatalyst becomes the main source

of **OH** and function as the main degrading species of RhB dye. Here **OH** and O<sub>2</sub><sup> $\text{−}$ </sup> radical help in the cleavage of conjugated structures (path b). On the other hand, RhB dye itself can absorb sunlight and goes to the excited state of dye (RhB<sup>\*</sup>). This excited dye can transfer electron to the CB of  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> to form RhB<sup>+</sup> as the LUMO of RhB and CB of  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> are -1.42 V and 0.33 V vs NHE [56, 57]. The RhB**+.** reacts with adsorbed dye to form deethylated products (path a) by the photo sensitization process (shown in Fig. 13). Again, the adsorbed molecular oxygen on the catalyst captures electron from the CB of  $Bi_2O_3$  to form  $O_2^{\bullet-}$ . The oxidant  $O_2^{\bullet-}$  radical reacts with adsorbed dye to degrade it. The holes in the HOMO react with adsorbed  $H_2O$  to form  $\bullet$ OH radical. It is worth mentioning that the photosensitization degradation of RhB occurs commonly via process involving N-deethylated step compared to photocatalytic process [58]. Therefore, de ethylated products are found in this study indicating that photosensitization process plays a vital role in the overall degradation of RhB although both path a and b might be operating simultaneously. In presence of Cl in the system,  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> is gradually converted to BiOCl in aqueous medium as observed in our previous study [12]. Although α or β Bi<sub>2</sub>O<sub>3</sub> comprise type 1 semiconductor junction, in situ generation of BiOCl makes the  $\alpha$  or β Bi<sub>2</sub>O<sub>3</sub> / BiOCl as type II junction and thus electrons are accumulated at the conduction band of β  $Bi<sub>2</sub>O<sub>3</sub>$ , holes are accumulated at the valence band of BiOCl, making effective separation of primary charge carriers. A plausible scheme stating the two paths are shown below.



In our system both nitrogen centre radical path 'a' and carbon centered radical path 'b' through initial formation of  $O_2$ <sup>--</sup> and •OH, occur simultaneously. But it seems that the reactions through path 'a' occur more effectively at the initial stage of first 90 minutes for which the N-deethylated products are the main intermediates. After 90 minute, destruction of chromophoric structure through path 'b' dominates over path a and finally after 150 minutes we get small compounds containing a phenyl backbone like phthalic acid (166), phthalic anhydride (148), dihydroxy benzoic acid (154) etc. Since de ethylated products are mainly formed after exposure of radiation for small time interval (60 min, 90 min) as obtained from mass and HPLC studies, it may be inferred that the breaking of ring structure is easy after de-ethylation.



**Fig. 13.** Plausible charge-transfer mechanism and photo catalytic degradation of RhB.

# **5. Conclusion**

We have successfully synthesized biphasic heterojunctions having various molar ratios of  $\alpha$  and  $β$  forms of Bi<sub>2</sub>O<sub>3</sub> by changing the concentration of bismuth nitrate precursor in glycerol solution following facile solvothermal method without calcination. The X-ray diffraction and microscopic studies reveal that the dimension of nanosphere and nanorod of two different forms produced are gradually increased with the initial concentration of the precursor. The optimum dosage of the synthesized catalyst, pH and  $H_2O_2$  concentration for better photocatalytic degradation of 10 ppm RhB are found to be 0.5  $gL^{-1}$ , 4 and 2 mM respectively. The efficiency of the degradation by the best  $\alpha$ -β Bi<sub>2</sub>O<sub>3</sub> composite is increased from 89% at pH 5 to 98% at pH 4 in 180 min under sunlight illumination. Addition of  $NO<sub>3</sub>$  and Cl salts increase the photocatalytic degradation of RhB by 12.1% and 9.0 %, while  $SO_4^2$  and  $CH_3COO^-$  salts decrease degradation in comparison to that found without salt addition. These results infer strong adsorption of  $NO<sub>3</sub>$  and Cl ions on the α-β Bi2O3 surface which seemingly facilates the greater generation of **.**OH radicals. The greater presence of N-deethylated products (DER, EER, DR, R) as confirmed from HRMS and HPLC techniques, indicates that the degradation mainly proceeds via photosensitization induced photocatalytic mechanism. In this process, RhB dye produce RhB**.+** and **.**OH radical which might be the main degrading agent. The decarboxylation and the ring cleavage of RhB were also observed on irradiation of sunlight for 150 min under optimized condition. The main nontoxic products are adipic acid, phthalic acid, 2- hydroxyl penta-di-ioic acid and 2, 5 di hydroxyl benzoic acid which are again mineralized to  $CO<sub>2</sub>$  and  $H<sub>2</sub>O$  on more exposure to sunlight.

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#### Reference

[1] E. Kevin, O. Dionysios, D. Dionysiou, Advanced oxidation processes for water treatment, J. Phys. Chem. Lett. 3 (2012) 2112−2113.

- [2] S.H.S. Chan, T.Y. Wu, J.C. Juan, C.Y. Teh, Recent developments of metal oxide semiconductors as photocatalysts in advanced oxidation processes (AOPs) for treatment of dye waste-water, J. Chem. Technol. Biotechnol. 86 (2011) 1130–1158.
- [3] J. Schneider, M. Matsuoka, M. Takeuchi, J. Zhang, Y. Horiuchi, M. Anpo, D.W. Bahnemann, Understanding  $TiO<sub>2</sub>$  photocatalysis: mechanisms and materials, Chem. Rev. 114 (2014) 9919−9986.
- [4] A. Mclaren, T. Valdes-solis, G.Q. Li, S.C. Tsang, Shape and size effects of ZnO nanocrystals on photocatalytic activity, J. Am. Chem. Soc. 131 (2009) 12540–1541.
- [5] M. Muruganandham, R. Amutha, G.J. Lee, S.H. Hsieh, J.J. Wu, M. Sillanpää, Facile fabrication of tunable  $\overline{Bi_2O_3}$  self-assembly and its visible light photocatalytic activity, J. Phys. Chem. C 116 (2012) 12906–12915.
- [6] F. Motahari, M.R. Mozdianfard, F. Soofivand, M. Salavati-Niasari, NiO nanostructures: synthesis, sharacterization and photocatalyst application in dye wastewater treatment, RSC Adv. 4 (2014) 27654–27660.
- [7] K. Nakata, A. Fujishima,  $TiO<sub>2</sub>$  photocatalysis: design and applications, J. Photochem. Photobiol. C Photochem. Rev. 13 (2012) 169–189.
- [8] L. Leontie, M. Caraman, M. Alexe, C. Harnagea, Structural and optical characteristics of bismuth oxide thin films, Surf. Sci. 507 (2002) 480–485.

[9] M. Drache, P. Roussel, J.P. Wignacourt, Structures and oxide mobility in Bi−Ln−O materials: heritage of  $Bi<sub>2</sub>O<sub>3</sub>$ , Chem. Rev. 107 (2007) 80–96.

[10] Y. Shi, L. Luo, Y. Zhang, Y. Chen, S. Wang, L. Li, Y. Long, F. Jiang, Synthesis and characterization of  $\alpha/\beta$ -Bi<sub>2</sub>O<sub>3</sub>with enhanced photocatalytic activity for 17 $\alpha$ -Ethynylestradiol, Ceram. Int. 43 (2017) 7627–7635.

[11] T.A. Gadhi, A. Hernández-Gordillo, M. Bizarro, P. Jagdale, A. Tagliaferro, S.E. Rodil, Efficient  $\alpha/\beta$ -Bi<sub>2</sub>O<sub>3</sub> composite for the sequential photodegradation of two-dyes mixture, Ceram. Int. 42 (2016) 13065–13073.

[12] K.K. Bera, R. Majumdar, M. Chakraborty, S.K. Bhattacharya, Phase control synthesis of α,  $β$  and  $α/β$  Bi<sub>2</sub>O<sub>3</sub> hetero- junction with enhanced and synergistic photocatalytic activity on degradation of a toxic dye, Rhodamine-B under natural sunlight, J. Hazard. Mater. 352 (2018) 182-191.

[13] R. Chen, Z.R. Shen, H. Wang, H.J. Zhou, Y.P. Liu, D.T. Ding, T.H. Chen, Fabrication of mesh-like bismuth oxide single crystalline nanoflakes and their visible light photocatalytic activity, J. Alloys Compd. 509 (2011) 2588–2596.

[14] X. Xiao, R. Hu, C. Liu, C. Xing, C. Qian, X. Zuo, J. Nan, L. Wang, Facile large-scale synthesis of  $\beta$ -Bi<sub>2</sub>O<sub>3</sub> nanospheres as a highly efficient photocatalyst for the degradation of acetaminophen under visible light irradiation, Appl. Catal. B Environ. 141 (2013) 433–443.

[15] Y. Qiu, M. Yang, H. Fan, Y. Zuo, Y. Shao, Y. Xu, X. Yang, S. Yang, Phase-transitions of α- and  $\beta$ -Bi<sub>2</sub>O<sub>3</sub>nanowires, Mater. Lett. 65 (2011) 780–782.

[16] J. Hou, C. Yang, Z. Wang, W. Zhou, S. Jiao, H. Zhu, In situ synthesis of  $\alpha-\beta$  phase heterojunction on  $Bi_2O_3$  nanowires with exceptional visible-light photocatalytic performance, Appl. Catal. B Environ. 142 (2013) 504–511.

[17] K. Brezesinski, R.Ostermann, P. Hartmansin, J. Brezesinski, Exceptional photocatalytic activity of ordered mesoporous  $β-Bi<sub>2</sub>O<sub>3</sub>$  thin films and electrospun nanofiber, Chem. Mater. 22 (2010) 3079–3085.

[18] S. Roy Chowdhury, S. Ghosh, S.K. Bhattachrya, Enhanced and synergistic catalysis of onepot synthesized palladium-nickel alloy nanoparticles for anodic oxidation of methanol in alkali, Electrochim. Acta 250 (2017) 124 –134.

[19] W. Ma, N. Wang, Y. Lu, Z. Lu, X. Tang, S. Li, Synthesis of magnetic biomass carbonbased  $Bi<sub>2</sub>O<sub>3</sub>$  photocatalyst and mechanism insight by a facile microwave and deposition method. New J. Chem. 43 (2019) 2888-2898.

[20] R. Ji, C. Ma, W. Ma Y. Liu, Z. Zhu, Y. Yan, Z-scheme  $MoS<sub>2</sub>/Bi<sub>2</sub>O<sub>3</sub>$  heterojunctions: enhanced photocatalytic degradation performance and mechanistic insight. New J. Chem. 4 3(2019) 11876-11886.

[21].F. He, J. wang, D. Deng, Effect of  $Bi_2O_3$  on structure and wetting studies of  $Bi_2O_3$ –ZnO– B2O3 glasses, J. Alloy. Compd. 509 (2011) 6332–6336.

[22] S. Wang, Y. Guan, L. Wang, W. Zhao, H. He, J. Xiao, S. Yang, C. Sun, Fabrication of a novel bifunctional material of  $BiO I/Ag_3VO_4$  with high adsorption–photocatalysis for efficient treatment of dye wastewater, Appl. Catal. B 168 (2015) 448–457.

[23] M. Popa, A. Totovana, L. Popescu, N. Dragan, M. Zaharescu, Reactivity of the Bi, Sr, Ca, Cu oxalate powders used in BSCCO preparation, J. Eur. Ceram. Soc. 18 (1998) 1265–1271. [24] Y. Sun, W. Wang, L. Zhang, Z. Zhang, Design and controllable synthesis of  $\alpha$ -/ $\gamma$ -Bi<sub>2</sub>O<sub>3</sub> homojunction with synergetic effect on photocatalytic activity. Chem. Eng. J. 211 (2012) 161- 167.

[25] A.H ezam, , K., Namratha, Q. A., Drmosh, Z. H., Yamani, K.Byrappa, Synthesis of heterostructured  $Bi_2O_3-CeO_2-ZnO$  photocatalyst with enhanced sunlight photocatalytic activity. Ceram. Int. 43(2017) 5292-5301.

[26] C. Wang, C. Shao, L. Wang, L. Zhang, X. Li, Y. Liu, Electrospinning preparation, characterization and photocatalytic properties of  $Bi<sub>2</sub>O<sub>3</sub>$  nanofibers, J. Colloid Interf. Sci. 333 (2009) 242–248.

[27] Y. Xiong, M. Wu, J. Ye, Q. Chen, Synthesis and luminescence properties of hand-like α- $Bi<sub>2</sub>O<sub>3</sub>$  microcrystals, Mater. Lett. 62 (2008) 1165–1168.

[28] H. Feng, H. Zijun, X. Junlin, L. Yuhui, IR and Raman spectra properties of  $Bi<sub>2</sub>O<sub>3</sub>$ -

ZnO-B<sub>2</sub>O<sub>3</sub>-BaO quaternary glass system, Am. J. Anal. Chem. 5 (2014) 1142–1150.

[29] Q. Huang, S. Zhang, C. Cai, B. Zhou,  $\beta$ - and  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> nanoparticles Synthesized via microwave-assisted method and their thotocatalytic activity towards the degradation of Rhodamine B, Mater. Lett. 65 (2011) 988–990.

[30] Y. Lu, Y. Zhao, J. Zhao, Y. Song, Z. Huang, F. Gao, Y. Li, Induced aqueous synthesis of metastable β-Bi<sub>2</sub>O<sub>3</sub>microcrystals for visible-light photocatalyst study, Cryst. Growth Des. 15 (2015) 1031–1042.

[31] L. Shan, G. Wang, D. Li, X. San, L. Liu, L. Dong, Z. Wu, Band alignment and enhanced photocatalytic activation of  $\alpha/\beta - Bi_2O_3$  heterojunctions via in situ phase transformation, Dalton Trans. 44 (2015) 7835-7843.

[32] R. Irmawati, M.N.N. Nasriah, Y.H. TaufigYap, S.B. A. Hamid, Characterization of bismuth oxide catalysts prepared from bismuth trinitrate pentahydrate: influence of bismuth concentration, Catal. Today 93 (2004) 701–709.

[33] H.E. Zhong, Y. Shaogui, J.U. Yongming, S. Cheng, Microwave photocatalytic degradation of Rhodamine B using TiO<sub>2</sub> supported on activated carbon: Mechanism implication. J. Environ. Sci. 21(2009) 268–272.

[34] F. Chen, J. Zhao, H. Hidaka, Highly selective deethylation of Rhodamine B: adsorption and photooxidation pathways of the dye on the  $TiO<sub>2</sub>/SiO<sub>2</sub>$  composite photocatalyst. Int. J. Photoenergy 5 (2003) 209–217.

[35] T. Selvamani, S. Anandan, L. Granone, D. W. Bahnemann, M. Ashokkumar, Phasecontrolled synthesis of bismuth oxide polymorphs for photocatalytic applications. *Mat. Chem. Front. 2* (2018) 1664-1673.

[36] C. Lops, A. Ancona, D. Cesare, K. Dumontel, B. Garino, N. Canavese, V. Cauda, Sonophotocatalytic degradation mechanisms of Rhodamine B dye via radicals generation by micro-and nano-particles of ZnO. Appl. Catal. B Environ. 243 (2019) 629-640.

[37]. S. Sood, A. Umar, S.K. Mehtaa, S.K. Kansal,  $\alpha$ -Bi<sub>2</sub>O<sub>3</sub> nanorods: an efficient sunlight active photocatalyst for degradation of Rhodamine B and 2,4,6-trichlorophenol. Ceram. Int. 41 (2015) 3355–3364.

[38] T. Kou, C. Jin, C. Zhang, J. Sun, Z. Zhang, Nanoporous Core-Shell Cu@Cu<sub>2</sub>O Nanocomposites with Superior Photocatalytic Properties towards the Degradation of Methyl Orange. RSC Adv. 33 (2012) 12636–12643.

[39] P. Mohammadyari, A. Nezamzadeh-Ejhieh, Supporting of mixed ZnS-NiS semiconductors

onto clinoptilolite nano-particles to improve its activity in photodegradation of 2 nitrotoluene, RSC Adv. 5 (2015) 75300–75310.

[40] H.R Pouretedal, A. Norozi, M.H. Keshavarz, A. Semnani, Nanoparticles of zinc sulfide doped with manganese, nickel and copper as nanophotocatalyst in the degradation of organic dyes. J. Hazard. Mater. 162 (2009) 674–681.

[41] F. H. AlHamedi, M. A Rauf, S. S. Ashraf, Degradation studies of Rhodamine B in the presence of UV/H<sub>2</sub>O<sub>2</sub>. Desalination 239 (2009) 159–166.

[42] B. Katherine, A. Benedict, S. McFall, C. Anastasio, Quantum Yield of Nitrite from the Photolysis of Aqueous Nitrate above 300 nm, Environ. Sci. Technol., 51 (2017) 4387−4395.

[43] W. Hua, F. Wub, W. Liua, J. Liuc, Nitrate-induced Photodegradation of Colorants and the Corresponding Mechanisms Study, J. Adv. Oxid. Technol. 21 (2018) 285-296.

[44] T.T. Nga Phan, A.N. Nikoloski, P.A. Bahri, D. Li, Adsorption and photo-Fenton catalytic degradation of organic dyes over crystalline  $LaFeO<sub>3</sub>$ -doped porous silica. RSC Adv. 8 (2018) 36181-36190.

[45] S. Azimia, A. Nezamzadeh-Ejhieh, Enhanced activity of clinoptilolite-supported hybridized PbS–CdS semiconductors for the photocatalytic degradation of a mixture of tetracycline and cephalexin aqueous solution, J. Mol. Catal. A: Chemical 408 (2015) 152–160

[46] P.V. Nidheesh, R. Gandhimathi, N.S. Sanjini, NaHCO<sub>3</sub> enhanced Rhodamine B removal from aqueous solution by graphite–graphite electro Fenton system, Sep. Purif. Technol. 132 (2014) 568–576

[47] A. Dhanya, K. Aparna, Synthesis and evaluation of  $TiO<sub>2</sub>/Chitosan$  based hydrogel for the adsorptional photocatalytic degradation of Azo and anthraquinone dye under UV light irradiation

#### Procedia Technol. 24 (2016) 611 – 618

[48] Factors influencing the photocatalytic degradation of Rhodamine B by  $TiO<sub>2</sub>$ -coated nonwoven paper, N. Barka, S. Qourzal, A. Assabbane, A. Nounah, Y. Ait-Ichou . J. Photochem. Photobiol. 195 (2008) 346–351

[49] B. Kraeutler, C.D. Jaeger. A.J. Bard, Direct observation of radical intermediates in the photo-Kolbe reaction-heterogeneous photocatalytic radical formation by electron spin resonance. J. Am. Chem. Soc. 100 (1978) 4903-4905.

[50] S. Sato*,*Photo-Kolbe Reaction at Gas-Solid Interfaces. J. Phys. Chem., 87 (1983) 3531- 3537.

[51] F. Chen, J. Zhao, H. Hidaka, Highly selective deethylation of Rhodamine B: adsorption and photooxidation pathways of the dye on the  $TiO<sub>2</sub>/SiO<sub>2</sub>$  composite photocatalyst, *Int.* J. Photoenergy 5 (2003) 209–217.

[52] Y. Zhang, J. Zhou, Biomaterials photodegradation pathway of Rhodamine B with novel Au nanorods @ ZnO microspheres driven by visible light irradiation, J. Mater. Sci. 53 (2018) 3149– 3162.

 [53] T. Watanabe, T. Takizawa, K. Honda, Photocatalysis through excitation of adsorbates. 1. Highly efficient Ndeethylation of rhodamine B adsorbed to cadmium sulfide, J. Phys. Chem. 81 (1977)1845–1851.

[54] T. Takizawa, T. Watanabe, K. Honda, Photocatalysis through excitation of adsorbates: A comparative study of Rhodamine B and methylene blue on cadmium sulfide, J. Phys. Chem. 82 (1978) 1391–1396.

[55] H. Fu, S. Zhang, Photocatalytic Degradation of RhB by Fluorinated  $Bi_2WO_6$  and Distributions of the Intermediate Products, Environ. Sci. Technol. 42 (2008) 2085–2091.

 [56] L. Pan, J. Zou, X. Liu, X. Liu, S. Wang, X. Zhang, L. Wang, Visible-light-induced photodegradation of Rhodamine B over hierarchical TiO<sub>2</sub>: effects of storage period and watermediated adsorption switch, Ind. Eng. Chem. Res. 51 (2012) 12782-12786.

[57] Y. Xu, M. Schoonen, The absolute energy positions of conduction and valence bands of selected semiconducting minerals, Am. Mineral. 85 (2000) 543-556

[58] P. Lei, C. Chen, J. Yang, W. Ma, J. Zhao, L. Zang, Degradation of dye pollutants by immobilized polyoxometalate with  $H_2O_2$  under visible-light irradiation. Environ. Sci. Technol.

39 (2005) 8466–8474.

#### **Declaration of interests**

 $\Box$  The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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